Geographic patterns of vertebrate diversity and identification of relevant areas for conservation in Europe


Abstract
Geographic patterns of vertebrate diversity and identification of relevant areas for conservation in Europe.— The 'EU Council conclusions on biodiversity post–2010' re–enforced Europe’s commitment to halt biodiversity loss by 2020. Identifying areas of high–value for biodiversity conservation is an important issue to meet this target. We investigated the geographic pattern of terrestrial vertebrate diversity status in Europe by assessing the species richness, rarity, vulnerability (according to IUCN criteria), and a combined index of the three former for the amphibians, reptiles, bird and mammals of this region. We also correlated the value of all indices with climate and human influence variables. Overall, clear geographic gradients of species diversity were found. The combined biodiversity index indicated that high–value biodiversity areas were mostly located in the Mediterranean basin and the highest vulnerability was found in the Iberian peninsula for most taxa. Across all indexes, the proportion of variance explained by climate and human influence factors was moderate to low. The results obtained in this study have the potential to provide valuable support for nature conservation policies in Europe and, consequently, might contribute to mitigate biodiversity decline in this region.

Key words: High–value biodiversity areas, Human influence, Richness, Rarity, Vulnerability.

Resumen
Patrones geográficos de diversidad de vertebrados e identificación de áreas relevantes para su conservación en Europa.— Las conclusiones del ‘Consejo de la UE sobre la biodiversidad post–2010’ reforzaron el compromiso europeo de detener la pérdida de la misma para el año 2020. La identificación de áreas de alto valor para la conservación de la biodiversidad resulta importante para alcanzar esta meta. En el presente estudio investigamos la distribución geográfica del estatus de la diversidad de vertebrados en Europa evaluando la riqueza de especies, rareza, vulnerabilidad (según criterios de la UICN) y un índice combinado de los tres anteriores para anfibios, reptiles, aves y mamíferos de esta región. Además, se correlacionó el valor de estos cuatro índices con variables climáticas e influencia humana. En general, se identificaron gradientes geográficos claros de diversidad de las especies. El índice combinado de biodiversidad indicó que, para la mayoría de los taxones, las áreas de alto valor de biodiversidad se encuentran principalmente en la cuenca mediterránea y la mayor vulnerabilidad en la península Ibérica. La proporción de variación explicada por el clima y la influencia humana fue de moderada a baja para todos los índices. Los resultados de este estudio tienen el potencial de proporcionar un valioso soporte científico para las políticas europeas de conservación de la naturaleza y, consecuentemente, pueden contribuir a mitigar la pérdida de biodiversidad en esta región.

Palabras clave: Áreas de alto valor de biodiversidad, Influencia humana, Riqueza, Rareza, Vulnerabilidad.

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**Introduction**

Assessing broad geographical patterns of species distribution is crucial to identify areas with highest species richness, rarity or vulnerability that are relevant for species conservation (Davies et al., 2006; Kati et al., 2004; Mittemeier, 2005; Myers et al., 2000; Ombe et al., 2005). Myers (1988) used the term ‘hotspots’ to refer to those areas with relevant biodiversity characteristics that are threatened with destruction. These areas usually harbour high species richness and a high number of endemic species (Myers et al., 2000). The identification of biodiversity hotspots has been mostly based on the amount of biodiversity per land unit area (Veech, 2000), although some efforts have also considered the distribution of biodiversity threats (Balmford et al., 2000; Fleishman et al., 2006; Rey Benayas & de la Montaña, 2003; Sierra et al., 2002).

Metrics that take biodiversity and the risk of species loss into account in a particular region are important for conservation efforts and allow the identification of areas that need urgent protection (Didier et al., 2010; Mangules & Pressey, 2000; Rey Benayas & de la Montaña, 2003). Identifying factors that affect species threats in a particular area may provide the bases for protection and inspire prevention measures to mitigate such threats and thus extinction risk. The relationships between human factors and biodiversity are important to assess such risk of extinction as human pressures are often related to large changes in biological diversity. However, the literature shows contradictory results. Previous studies report that human influence may affect species’ spatial distribution both negatively and positively (Young et al., 2005).

On the one hand, human factors, such as human activities (Araújo et al., 2002; Cincotta et al., 2000; Clergeau et al., 2006; Donald et al., 2001) and, in particular, the alteration of habitats (Kiesecker et al., 2001; Peres et al., 2010) are major causes of biodiversity loss (Brooks et al., 2002; Cardillo et al., 2004; Gaston, 2006; McKee et al., 2003; McKinney, 2001; Singh, 2002; Van Rensburg et al., 2004). On the other hand, several studies have even shown a positive relationship between human density and biodiversity, indicating that species–rich areas and human settlements often co–occur (Albuquerque & Rueda, 2010; Luck, 2007; Maffi, 2005; Sutherland 2003). However, this might be a purely correlative effect in many instances, particularly for species that are associated with farming and human habitation such as aphids (Pautasso & Powel, 2009) or ants (Schlick–Steiner et al., 2008) that may behave as invasive pests causing an absolute loss of diversity by displacing other species.

The present study joins previous conservation biogeography efforts to identify critical areas to protect European vertebrate diversity (Araújo & Pearson, 2005; Jelaska et al., 2010); it aimed to document geographic patterns of species richness, rarity, vulnerability, and a combined index of the three former measures at the 50–km grain resolution for each major taxa. We also analyzed relationships between human influence and these biodiversity indices, highlighting key areas for vertebrate conservation. Our analysis provides insights into how to address anthropogenically–derived conservation issues.

**Material and methods**

**Distribution data**

Distribution data from atlas maps for amphibian, reptile, bird, and mammal species in Europe were obtained from Gasc et al. (1997); Hagemeijer & Blair (1997) and Mitchell–Jones et al. (1999). These maps were digitalized and processed in Arc GIS 9.3 in a grid comprising 2,194 UTM cells of 50 x 50 km each. All islands, except Great Britain, and cells with less than 50% land cover were excluded from the analyses. Preliminary data analyses identified some cells with abnormally low amphibian and reptile richness compared with nearby cells. We identified these cells as outliers and they were excluded from analysis.

**Criteria for identifying areas of high–value diversity**

We followed Rey Benayas & de la Montaña (2003) to identify areas of high–value diversity of the various taxonomic groups. The following biodiversity criteria were assessed in all cells: a) species richness, b) rarity, c) vulnerability, and d) a combined index of biodiversity that integrates the three former criteria.

Rarity (R) was computed for each cell r as:

\[ R = \sum_{i=1}^{n} \left( \frac{1}{n_i} \right) / S_r \]

where \( n \) is the number of cells in which species \( i \) is present, and \( S_r \) is the cell’s species richness.

For vulnerability (V), we first ranked the five threat categories defined by the International Union for Nature Conservation (IUCN, 2006) as: (1) non–threatened, (2) insufficiently known, (3) rare, (4) undetermined or vulnerable, and (5) endangered, and then computed the index for each cell as:

\[ V = \sum_{i=1}^{s} \left( \frac{v_i}{S_r} \right) \]

where \( v_i \) is the vulnerability rank of species \( i \), and \( S_r \) is the richness of cell \( r \). Initially, we also computed this index using the similar categories defined by the European Nature Information System (EUNIS, 2005) but obtained similar results (not shown) which led us to omit this index from the study.

Then, we calculated the combined index of biodiversity (C), which jointly evaluates the species richness, rarity and vulnerability for each cell:

\[ C = \sum_{i=1}^{s} \left( \frac{1}{n_i} \right) v_{ir} \]

in which species richness is implicit in the expression \( \sum_{i=1}^{s} \frac{1}{n_i} \) and vulnerability by \( v_{ir} \).

Finally, we calculated a standardized biodiversity index (SBI) by dividing the combined index of biodiversity of each taxonomic group in every cell by its
mean across all cells. Next, we summed the four standardized combined indices. The \( SBI \) formula is:

\[
SBI = \sum_{j=1}^{4} \frac{m_j}{S} \sum_{i=1}^{n_j} \left( V_{ji} \right)
\]

where \( m_j \) refers to the mean combined index of biodiversity of the taxonomic group \( j \) across cells.

Climate and human influence variables

We generated 21 variables to explain geographic patterns of vertebrate richness, rarity and vulnerability. These comprised the 19 climate variables of the WorldClim database (annual mean temperature, mean diurnal range, isothermality, temperature seasonality, maximum temperature of warmest month, minimum temperature of coldest month, temperature annual range, mean temperature of wettest quarter, mean temperature of driest quarter, mean temperature of warmest quarter and mean temperature of coldest quarter, annual precipitation, precipitation of wettest month, precipitation of driest month, precipitation seasonality, precipitation of wettest quarter, precipitation of driest quarter, precipitation of warmest quarter, and precipitation of coldest quarter; Hijmans et al., 2005), and two surrogates of human influence, namely human population density and a habitat fragmentation index. Human density was obtained from the Gridded Population of the World [urban mapping project, version 3 produced by the Center for International Earth Science Information Network (CIESIN) and available at: http://sedac.ciesin.columbia.edu/gpw/ (last accessed February 2012)]. The habitat fragmentation index measures the fragmentation of land by urbanization, transport infrastructure and agriculture. It calculates how many natural complexes are found within each cell and the compactness of these complexes (average size of complex in a cell versus total area of complexes in the cell). This index was produced by the European Environment Agency and is available at http://www.eea.europa.eu/data-and-maps/figures/fragmentation–by–urbanisation–infrastructure–and–agriculture (last accessed February 2012).

Data analysis

Initially, relationships among the four biodiversity variables (species richness, rarity, vulnerability and the combined index of biodiversity) within taxonomic groups were examined by means of Spearman rank correlation using Bonferroni correction for multiple comparisons. We also performed a principal component analysis (PCA) including all biodiversity variables (species richness, rarity, vulnerability, and the combined index of biodiversity) for each taxonomic group as well as the combined biodiversity index to highlight relationships among multiple and highly correlated variables. Additionally, relationships of each biodiversity index with climate and human influence variables were investigated by means of a redundancy analysis–based variation partitioning (Borcard et al., 1992; Legendre & Legendre, 1998; Péres–Neto et al., 2006). This analysis provides a synthetic view of the relationships by partitioning the variation of a response variable in the study area (i.e. a biodiversity index of a particular vertebrate group) into components independently and jointly explained by groups of explanatory variables (i.e. climate variables and human factors in this study). Finally, we also took into account the results of Whittaker et al. (2007) who found that relationships of amphibian, bird, and mammal (but not reptile) species richness with solar radiation (a measure of the amount of energy available in the environment) shifted from positive in northern Europe to negative in the south of this region, and that the line separating these two zones was different for each group. Thus, we repeated the above–mentioned analyses separately for each of these regions and species groups. All analyses were performed in R (R Development Core Team, 2009) using the ‘vegan’ package (Oksanen et al., 2009).

Results

Geographical patterns of vertebrate diversity

There are 817 terrestrial vertebrate species in our study area, of which 52 are amphibians, 108 reptiles, 515 birds, and 142 mammals. Except for birds, which showed higher species richness in central European regions, there was a tendency of the richness of the other three vertebrate groups to increase southwards, with picks of highest richness values occurring in central Europe for amphibians and mammals, and in Mediterranean areas (Iberian peninsula and Greece) for reptiles (figs. 1A–1D). The overall geographic pattern of rarity (\( R \)) was similar for the four taxonomic groups, with rarity generally increasing southwards, although for birds and mammals it also showed secondary peaks in the north (Norway, Sweden and Finland; figs. 1E–1H).

Higher values of the vulnerability index (\( V \)) based on the IUCN threat categories for amphibians were recorded in northern Europe to south–eastern Portugal and west–central Spain; for reptiles in France and Germany primarily, and Norway, Sweden and Romania secondarily; and for birds and mammals across the Iberian Peninsula, Poland, Ukraine and Romania, with mammals also picking in north–eastern Europe (figs. 1I–1L).

Amphibians and reptiles showed a clear north–to–south gradient of increasing values of the combined index of biodiversity (\( C \)), mammals did the same albeit, with a more patchy distribution, and birds showed no clear trend, with high values occurring in localized areas of southern (Iberian and Greek peninsulas), central (e.g. Great Britain and Hungary) and northern (Norway, Sweden and Finland) Europe (figs. 1M–1P).

Highest values of the standardized biodiversity index (\( SBI \)) that integrates all biodiversity criteria for the four taxonomic groups were mainly observed in the Mediterranean basin, especially in Portugal, Spain, Greece and Bulgaria, with a secondary peak in Northern Europe (fig. 2).
Relationships among biodiversity criteria

Correlation analyses between species richness, rarity, vulnerability, and the combined biodiversity index within each vertebrate group indicated that almost all these biodiversity estimates were significantly and positively correlated (table 1). The combined biodiversity index was positively correlated with all estimates and especially with rarity for all groups.

The two first axes of the PCA performed on all biodiversity criteria absorbed 36.8% and 18.1% of the variation, respectively. The visual inspection of this graph revealed association of rarity, the combined index and the standardized biodiversity index on one side, and of vulnerability and species richness on the other side (fig. 3). Taxonomic groups were spread throughout the PCA bi-plot; however, it is noticeable the fact that the bird diversity criteria are relatively independent from those of all remaining taxa (fig. 3).

Variation of vertebrate diversity explained by climate and human influence

The proportion of variation explained by climate and human influence variables was highest for richness, especially for amphibians (41%) and reptiles (42%) (table 2). Rarity, vulnerability and combined biodiversity indexes were, in general, less associated with climate and human influence variables. In all cases, climate contributed more than human influences to explain these biodiversity variables. This was also reflected in the results of the variation partitioning analyses conducted separately for north and south Europe for amphibians, birds and mammals, although more variation was explained by the models for the north (table 2).

Discussion

This study identified high-value diversity areas for amphibians, reptiles, birds, and mammals in Europe by documenting the geographic distribution of five biodiversity criteria and analysing their relationships with climate and human influence factors. For most groups (amphibians, reptiles and mammals) we observed a general north-to-south gradient of increasing richness, whereas for birds, the patterns were more complex and richness picked at central European regions. Still, climate was more important than human influences in driving the patterns in all cases. Similar richness gradients and relationships with climate have been reported by previous studies for these taxa across Europe (Araújo & Pearson, 2005; Carrascal & Díaz, 2003; Nogués-Bravo & Martínez-Rica, 2004; Olalla-Tárraga et al, 2006; Qian & Xiao, 2012; Rodríguez et al, 2005; Rojas et al., 2001).

We also found a strong and positive correlation between rarity and the combined index of biodiversity for all groups, which highlights rarity as a key criterion to identify high-value biodiversity areas over broad geographical extents. This supports
previous claims pointing out that rarity is likely to be more effective than richness to identify priority areas for conservation (Williams et al., 1996). This result is important, since richness is the conservation criterion that is used by decision makers most often (Médail & Quézel, 1997; Reyers et al., 2000; Rodrigues et al., 2004).

In general, for the four biodiversity criteria analysed, the proportion of variation explained by climate and human influence factors was moderate to low, suggesting that other factors might be important for the described geographical pattern of vertebrate diversity in Europe. Thus, the patterns found for amphibians and reptiles may be related to the lower dispersal capacity of these groups compared to that exhibited by other vertebrates, as species with low dispersal rates need a longer time to colonize sites away from their origin (Aragón et al., 2010; Araújo & Pearson, 2005), which in turn might be associated with their higher levels of endemism (Williams et al., 2000). In agreement, Araújo & Pearson (2005) reported low levels of equilibrium (i.e. the time needed to reach saturated communities) between the distribution of reptile and amphibian species in Europe and current climate, whereas they found that major ice-age refugia (Iberia, Italy and the Balkans) were key determinants of the current distributions of these species across this region (see also Whittaker et al., 2007). The contrasted geographical patterns found for bird richness in Europe (see fig. 1C), and the relatively independent location of this taxon with respect to all other taxa in the ordination of biodiversity criteria (see figure 3), may be related with the location of speciation centres, dispersal capacity and environmental preferences of the species of this taxon (Covas & Blondel, 1998). Also, bird and mammals appear to have been under a strong selective pressure by human disturbance in the northern hemisphere since the last glaciation, which may have also played a relevant role in driving the diversity patterns of these groups (Nogués–Bravo & Matrínez–Rica, 2004; Walther et al., 2002). Previous results have indicated that areas with high species rarity and vulnerability are usually associated with

Table 1. Spearman rank correlation coefficients between criteria used to identify areas of high–value diversity within taxonomic groups in Europe. Coefficients in bold are significant at \( p < 0.05 \) after applying Bonferroni corrections for multiple comparisons: S. Richness; R. Rarity; V. Vulnerability; C. Combined index.

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Tabla 1. Coeficientes de correlación de rango de Spearman entre los criterios utilizados en la identificación de las áreas de gran valor en cuanto a diversidad de los grupos taxonómicos de Europa. Los coeficientes en negrita son significativos para \( p < 0.05 \) tras aplicar las correcciones de Bonferroni para comparaciones múltiples: S. Riqueza; R. Rareza; V. Vulnerabilidad; C. Índice combinado.
low habitat variety, forest loss, human impacts and climate change (Carrascal & Palomino, 2006; Mainka & Howard, 2010; Nuñeza et al., 2010; Vié et al., 2009), in agreement with theory and empirical evidence that relate population declines with disturbance and habitat homogenization (Echeverrería et al., 2004; Rey Benayas et al., 1999). However, our results show a weak association between rarity and vulnerability of these taxa with climate and human influence variables. This difference may be related to the coarser grain used in this study, in agreement with suggestions that the relationship between the ecological characteristics of a given species and its rarity and vulnerability value are scale–dependent (Murray & Lepschi, 2004). Even though our results suggest relatively minor effects of climate and human influence variables on vertebrate rarity and vulnerability, it should be noted that these results were obtained for a particular geographical extent (Europe) and grain (cells of 50 km²), and we cannot discard a stronger role of human influence at smaller scales (e.g. see Derraik & Phillips, 2010; Nuñeza et al., 2010; Rowley et al., 2010). Additionally, the IUCN Red List clearly shows that many vertebrate species are under threat of extinction mainly as a direct or indirect result of human activities and climate change (Vié et al., 2009).

This study identified the Mediterranean basin as one of the richest, rarest and most vulnerable areas of Europe in terms of vertebrate diversity, and supports the tenet that Mediterranean basin biodiversity is under strong threat (see fig. 1M–1P).
This agrees with the findings of Myers et al. (2000), who recognized the Mediterranean as one of the 25 Global Biodiversity Hotspots. These threats are often attributed to human disturbance, natural disasters, habitat loss and degradation, pollution, or invasive alien species (Vié et al., 2009). However, human influence factors explained a small proportion of the variance of each of the four biodiversity criteria that we investigated. Further research might establish to what extent detection of human influence on diversity patterns are dependent on grain in studies conducted in large areas. Irrespective, our data allow us to conclude that using a range of biodiversity criteria is necessary to accurately identify high-value diversity areas on a large geographic scale.

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References


